Author's personal copy

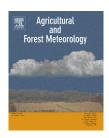
AGRICULTURAL AND FOREST METEOROLOGY 149 (2009) 373-384



available at www.sciencedirect.com



journal homepage: www.elsevier.com/locate/agrformet



Throughfall in an evergreen-dominated forest stand in northern Thailand: Comparison of mobile and stationary methods

Alan D. Ziegler ^{a,*}, Thomas W. Giambelluca ^b, Mike A. Nullet ^b, Ross A. Sutherland ^b, Chatchai Tantasarin ^c, John B. Vogler ^d, Junjiro N. Negishi ^e

ARTICLE INFO

Article history: Received 16 July 2007 Received in revised form 25 August 2008 Accepted 11 September 2008

Keywords:
Interception loss
PKEW
Spatial and temporal throughfall distribution
Tropical SE Asia
Wind

ABSTRACT

Throughfall determined by stationary and mobile methods in a disturbed evergreen-dominated forest stand in northern Thailand was 82% of rainfall (1134 mm) during a 4-month study period in the monsoon rain season of 2002. Associated coefficients of variation and standard errors were \leq 10% and 2%, respectively, for both methods. Agreement between four stationary trough collectors and 20 mobile standard gauge collectors was achieved only after 35 sampling occasions, having a total rainfall depth >700 mm, and included one storm event >100 mm. Several canopy trees contributed to points with throughfall > rainfall by channeling stemflow to common drip points on the trunk and large limbs. However, no significant correlation was observed between throughfall point measurements and corresponding canopy cover. Although 180-point measurements of throughfall provided a realistic representation of the spatial variability within the 500-m² forest stand, it is questionable that they duplicated the basin-scale variability, which would be affected both by tree gaps and variable topographically related rain shadow effects.

© 2008 Elsevier B.V. All rights reserved.

Introduction

Interception loss is the portion of the rain captured by leaves, stems, and branches of a vegetative cover that is subsequently evaporated (Hewlett, 1982; Dingman, 1994). Throughfall is the portion falling directly through and/or dripping to the ground from the canopy. Stemflow is the portion that reaches the ground by flowing down the trunks. Throughfall and stemflow are commonly measured as a

means of estimating interception loss (Bruijnzeel, 2000; Dunkerley, 2000). Of the two, throughfall is typically given more attention because the stemflow component is generally small: <5–10% of rainfall (Bruijnzeel, 1990, 2000; Levia and Frost, 2003). Throughfall quantification is important for other types of investigations, including the study of solute inputs to forest ecosystems, runoff generation, soil erosion below canopies, and distribution of soil moisture on forest floors (e.g., Bouten et al., 1992; Robson et al., 1994; Calder,

E-mail address: adz@hawaii.edu (A.D. Ziegler). 0168-1923/\$ – see front matter © 2008 Elsevier B.V. All rights reserved. doi:10.1016/j.agrformet.2008.09.002

^a Geography, National University Singapore, Singapore

^b Geography, University of Hawai'i Mānoa, Honolulu, HI, USA

^c Faculty of Conservation, Kasetsart University, Bangkok, Thailand

^d Program on Environment, East-West Center, Honolulu, HI, USA

^e Forestry and Forest Product Research Institute, Tohoku Branch, Japan

^{*} Corresponding author.

2001; Schroth et al., 2001; Raat et al., 2002; Hölscher et al., 2003; Schume et al., 2003).

Field experiments have demonstrated substantial spatial and temporal variability in throughfall at sites around the world (Levia and Frost, 2006). Variability is affected by season (e.g., Stout and McMahon, 1961; Peterson and Rolfe, 1979; Staelens et al., 2006; Cuartas et al., 2007), rainfall characteristics (e.g., Huber and Iroumé, 2001; Raat et al., 2002; Hall, 2003; Carlyle-Moses et al., 2004), and forest stand characteristics, including canopy structure, density, type, and level of disturbance (e.g., Helvey and Patric, 1965; Stogsdill et al., 1989; Asdak et al., 1998; Hall and Roberts, 1990; Aboal et al., 2000; Crockford and Richardson, 2000; Chappell et al., 2001; Loescher et al., 2002; Germer et al., 2006; Levia and Frost, 2006). Confounding the comparison of throughfall among sites is the uncertainty associated with a particular type of measurement method employed and the number of gauges involved in the estimate (e.g., Helvey and Patric, 1965; Czarnowski and Olszewski, 1970; Calder and Rosier, 1976; Bruijnzeel and Wiersum, 1987; Lloyd and Marques, 1988; Kostelnik et al., 1989; Forti and Neal, 1992; Rodrigo and Avila, 2001; Holwerda et al., 2006).

This study quantified throughfall in an evergreen-dominated forest stand in tropical northern Thailand using both a stationary and a mobile collection method. Objectives were: (1) calculate throughfall percentage for the stand; (2) determine whether stationary and mobile measurement techniques produce similar throughfall estimates; and (3) identify the spatial throughfall distribution.

2. Study site

Research was conducted in the Pang Khum Experimental Watershed (PKEW) in northern Thailand (19°03′N, 98°39′E). Pang Khum is located on the boundary of Samoeng and Mae Taeng Districts in Chiang Mai Province, approximately 60 km NNW of Chiang Mai (Ziegler et al., 2004). The 93.7-ha PKEW (Fig. 1a) is a headwater drainage of the Ping River, a major tributary to the Chao Phraya River. Elevation ranges from 1100 to 1600 m. The area has a monsoon rainy season extending from mid-May to October, which accounts for 80–90% of the annual rainfall total (1200–2000 mm).

3. Methods

3.1. Experimental setup

The study was conducted from 21 June to 7 October 2002 in a $500 \, \mathrm{m}^2$ study plot situated within a disturbed evergreendominated montane forest stand (Fig. 1; Tables 1 and 2). The plot was divided into $20.5 \, \mathrm{m} \times 5 \, \mathrm{m}$ quadrants (Fig. 1e). Throughfall was recorded using four stationary gauges and 20 standard-type, mobile throughfall gauges (rovers) that were relocated to unique locations during each of 11 periods (Table 1; Fig. 1). Periods contained 2–6 sampling occasions, for a total of 49 during the study (Table 1). A sampling occasion refers to the interval over which throughfall was collected; each occasion included at least one distinct rainfall event.

Rainfall was measured at meteorological tower #401, located approximately 70 m from the study plot (Fig. 1a). Rainfall data were recorded at 1-min intervals at a height of 22 m – approximately 1 m above the canopy – with a 0.153-m diameter Texas Instruments tipping-bucket rain gauge and Campbell Scientific 21x data logger. Wind speed was monitored with a Met-One anemometer at the same height on the 401 tower.

3.2. Throughfall measurements

Stationary measurement gauges consisted of fabricated galvanized steel tipping-bucket gauges that received throughfall water captured in troughs that extended under the forest canopy (Fig. 2a). The tipping mechanism is a larger version of a commercial tipping-bucket rain gauge with a solid state reed switch that is monitored by a Campbell CR10x data logger (Fig. 2b). The initial volume of throughfall required to produce one tip was 150 cm³ (0.2 mm). A dynamic calibration correction was then applied to account for differences in tip volume over the range of observed tipping rates (Calder and Kidd, 1978; Marsalek, 1981; Humphrey et al., 1997). Each collection trough was 43 mm wide, and had a triangular-shaped channel (120° angle) and 25 mm vertical risers to reduce rain splash loss (Fig. 2b). Four stationary gauges were installed to sample a variety of dense and sparse portions of the canopy (Fig. 1d). Each of three troughs was 6 m long; therefore, the sampling area for each gauge was approximately 0.77 m² (after correcting for trough angle). A total of 3.1 m² was sampled by the four stationary gauges. Troughs were positioned 0.5–1.0 m above the ground to limit interference by understory vegetation.

The rovers were galvanized funnels (angle = 60° ; radius = 0.130 m; area = 0.054 m²) that drained into 5-L collection bottles (Fig. 2c). Unlike the 401 rainfall gauge, the funnels did not have vertical sidewalls or screens to prevent splash loss. During each of the 11 measurement periods, 5–20 rovers were positioned systematically at unique locations such that throughfall was measured at 12–13 locations in each of the 20 quadrants (Table 1, Fig. 1f). Total area sampled at 180 unique measurement locations was 9.7 m². Minimum rover separation distance was 2 m, while the maximum separation distance was 33 m. Throughfall volumes were measured daily, but exceptions were made if researchers were away for 2–3 days. Evaporation losses from the collection bottles were assumed to be negligible.

Splash loss errors were corrected by establishing regression equations with depths measured by the tipping-bucket rain gauge during several storms prior to experimentation. Throughfall was calculated by three methods: (1) mean of throughfall estimates determined for 49 sampling occasions (stationary method); (2) mean of 20 rover throughfall estimates determined for 49 sampling occasions (rover method); and (3) mean of 180 unique rover-based throughfall values, determined for two to six sampling occasions (point method). Methods 1 and 2 were used for estimating stand throughfall and comparing stationary and mobile methods (objectives 1 and 2). The point method allowed investigation of the spatial distribution of throughfall (objective 3).

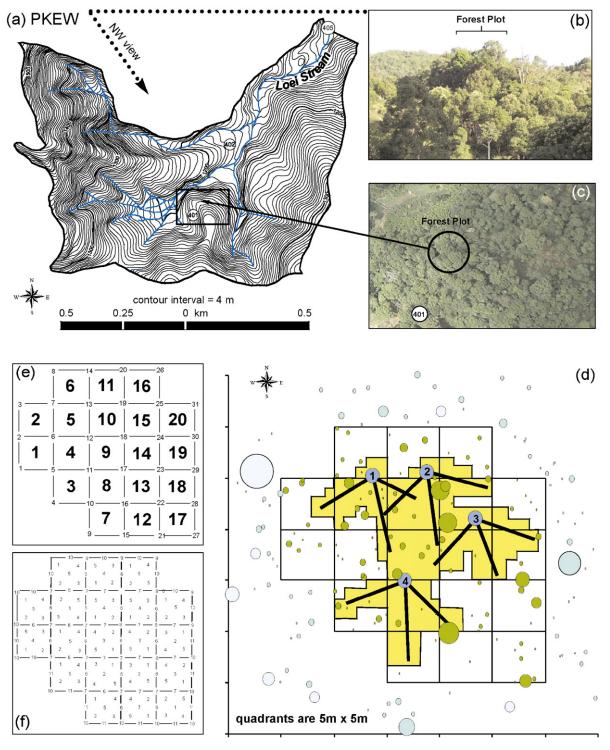


Fig. 1 – (a) Location of experimental forest plot, station 401 (rainfall and wind), and station 402 (above canopy PAR) in the Pang Khum Experimental Watershed (PKEW) in northern Thailand. (b) The experimental forest stand viewed from the northwest; the approximate location of the forest study plot is indicated. (c) Forest stand from light aircraft (500 m altitude). (d) Arrangement of four stationary throughfall collectors (numbered with three collection troughs) and location of 102 trees within the 500 m² experimental plot. Circles are tree trunks exaggerated in scale by 200%. Only trees within about 5 m outside the plot are also shown. The length of the stationary gauge collection troughs are to scale. Shading represents the subplot corresponding to the area roughly sampled by the stationary gauges (200 m²). (e) Twenty 5 m \times 5 m quadrants comprising the experimental forest study plot (large numbers). The small numbers represent locations where LAI values were measured in addition to the center of each of 20 quadrants (51 total locations). (f) Locations of 180 throughfall point measurements made with the mobile rovers. The numbers refer to one of 11 periods during which measurements were conducted (Table 1).

Table 1 – Setup of rovers and stationary gauges for 11 measurement periods in 2002.								
Period	Start day	Stop day	Period duration (days)	Sampling occasions in period	Number of rovers deployed	Total rainfall (mm)	Method 1 stationary TF (%) ^a	Method 2 rover TF (%)ª
1	21 June	28 June	7	6	20	76	66	86
2	28 June	8 July	10	4	20	74	70	71
3	8 July	21 July	13	6	20	76	65	65
4	21 July	4 Aug	14	5	20	93	72	72
5	4 Aug	12 Aug	8	5	20	59	76	75
6	12 Aug	23 Aug	11	4	20	101	81	79
7	23 Aug	28 Aug	5	5	20	112	85	95
8	28 Aug	9 Sept	12	4	6	154	97	88
9	9 Sept	12 Sept	3	4	5	48	84	86
10	12 Sept	24 Sept	12	4	20	177	89	90
11	24 Sept	7 Oct	13	2	9	164	86	86
Total			108	49	180 ^b	1134	82 ^c	82 ^c

^a Throughfall (percent of rainfall depth) values are means of four stationary or 5–20 mobile (rover) gauges for each sample period.

3.3. Statistical handling

To assess statistical differences among the rover and stationary methods, we used parametric approaches that allowed paired testing: e.g., paired t-test and ANCOVA (with

Table 2 – Name, type (evergreen or deciduous), and number of tree species (102 total) in the 500 m^2 forest patch.

Name	Туре	No.
Wendlandia tinctoria (Roxb.) DC. (Rubiaceae)	Е	27
Stryax benzoides Craib (Styracaceae)	E	15
Castanopsis tribuloides (Sm.) A. DC. (Fagaceae) ^a	E	12
Lithocarpus elegans (Bl.) Hatus. ex Soep. (Fagaceae) ^a	E	7
Castonopsis diversifolia (Kurz) King ex Hk. f. (Fagaceae) ^a	E	6
Helicia nilagirica Bedd. (Proteaceae) ^a	Е	5
Heliciopsis terminalis (Kurz) Sleum. (Proteaeceae) ^a	Е	5
Machilus bombycina King ex Hk. f. (Lauraceae) ^a	Е	3
Aporusa villosa (Lindl.) Baill. (Euphorbiaceae)	D	2
Camellia pleurocarpa (Gagnep.) Sealy (Theaceae)a	E	2
Engelhardia serrata Bl. var. serrata (Juglandaceae) ^a	D	2
Falconeria insignis Roy. (Euphorbiaceae) ^a	D	2
Antidesma bunius (L.) Spreng. (Euphorbiaceae) ^a	E	1
Artocarpus lanceolata Trec. (Moraceae) ^a	E	1
Cinnamomum iners Rienw. ex Bl. (Lauraceae) ^a	E	1
Dillenia parviflora Griff.var. kerrii (Craib)	D	1
Hoogl. (Dilleniaceae)		
Eugenia albiflora Duth. ex Kurz (Myrtaceae) ^a	E	1
Euonymus cochinchinensis Pierre (Celastraceae) ^a	E	1
Garcinia cowa Roxb. (Guttiferae) ^a	E	1
Kydia calycina Roxb. (Malvaceae)	D	1
Litsea salicifolia Nees ex Roxb. (Lauraceae) ^a	E	1
Olea rosea Craib (Oleaceae) ^a	E	1
Rhus chinensis Mill. (Anacardiaceae)	D	1
Saurauia roxburghii Wall. (Saurauiaceae) ^a	E	1
Symplocos cochinchinensis (Lour.) S. Moore ssp.	E	1
laurina (Retz.) Noot. (Symplocaceae) ^a		
Ternstroemia gymnanthera (Wight & Arn.) Bedd. (Theaceae) ^a	Е	1
a Primary growth species		

^a Primary growth species.

rainfall as the interaction term). The minimum number of gauges (N_{min}) required to estimate throughfall based on the spatial coefficient of variation (CV = mean/std dev × 100%) for a desired error (ε) and confidence level was estimated as follows (cf. Kimmins, 1973):

$$N_{\min} = \frac{z_c^2 \times CV^2}{\epsilon^2} \tag{1}$$

where z_c is the critical value at the stated confidence interval. Herein, we use $\varepsilon = 10\%$ and $z_c = 1.96$ for the 95% confidence interval (cf. Carlyle-Moses et al., 2004; Holwerda et al., 2006).

Variograms were used to examine spatial dependence between throughfall point measurements (Davis, 2002). Semivariance (γ_h) was calculated as follows:

$$\gamma_{h} = \frac{\sum (x_{i} - x_{i+h})^{2}}{2n} \tag{2}$$

where x_i is throughfall at any point measurement location, x_{i+h} is throughfall at another location that is distance h away (lag); and n is the number of data pairs that are distance h apart.

Prior to all statistical testing, skewness (g) was quantified to assess normality (Zar, 1999). When values of g exceeded critical values for two-tailed testing at α = 0.05, a square-root transformation was applied (cf. Lloyd and Marques, 1988; Holwerda et al., 2006).

3.4. Leaf area index (LAI) and canopy density

Below-canopy photosynthetically active radiation (PAR) measurements were taken in the center of each of the 20 quadrants and at the 31 quadrant intersection/boundary points (Fig. 1f). PAR was measured with a Decagon AccuPAR LAI/PAR Ceptometer. LAI was calculated using the Decagon inversion model (Decagon Services, 2004), which takes into account above- and below-canopy PAR, canopy structure, and solar elevation angle (based on time, day, and latitude). Above-canopy PAR was measured simultaneously with a separate PAR sensor at station 402 (Fig. 1a).

^b A total of 835 rover measurements were taken at 180 unique locations over the course of 11 periods (seven measurements during period 1 were discarded because of experimental error). Four stationary gauges were deployed in the same location throughout the study (see Fig. 1). ^c Total values are means of either 20 rovers or four stationary gauges, calculated for total throughfall (i.e., not column means).

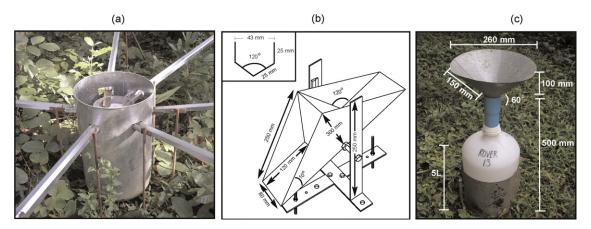


Fig. 2 – (a) Stationary collector with six collection troughs (only three troughs were used in this study because site was on non-level terrain). (b) Schematic of stationary gauge tipping-bucket mechanism; the inset shows the dimensions for the collection troughs. (c) Dimensions of the mobile rover throughfall collector; the collection bottle rests in a steel ring that was anchored to the ground to level/secure the funnel.

At 180 throughfall measurement locations (including all 51 LAI measurements points, Fig. 1e), a canopy density index (CDI) was determined as the product of canopy cover and canopy depth. Cover was quantified by comparing the overhead canopy cover with values on pattern cards ($\pm 10\%$ estimated accuracy). Depth was determined both with a measuring tape and extension rod (for heights <4 m) and via trigonometric methods (heights >4 m). The CDI has the same units as LAI (m² m $^{-2}$).

4. Results

4.1. LAI

Leaf area index estimated at 51 locations ranged from 0.2 to $4.1 \, \text{m}^2 \, \text{m}^{-2}$. CDI values determined at the same 51 locations ranged from 0.6 to $7.6 \, \text{m}^2 \, \text{m}^{-2}$. The following logarithmic regression was established between these two variables ($r_{\text{adi}}^2 = 0.65$; p-value < 0.0001; Fig. 3a):

$$LAI = 1.062 + 1.304 \ln CDI$$
 (3)

Eq. (3) was then used to estimate LAI for the remaining 129 throughfall point measurement locations. LAI was interpolated throughout the plot using an inverse distance weighting function (Fig. 4a). Most high LAI values were associated with a few large canopy trees, whereas many low LAI values were grouped in the SW corner, where tree density immediately outside the forest plot was low.

4.2. Rainfall

During the 108-day experiment, 1134 mm of rainfall was recorded (Table 1). The smallest rainfall depth during any 1 of the 11 measurement periods was 48 mm; the largest was 177 mm (Table 1). The smallest depth measured during any of the 49 sampling occasions was 0.88 mm; the maximum was 159 mm; and the median was 12.6 mm. During three sampling

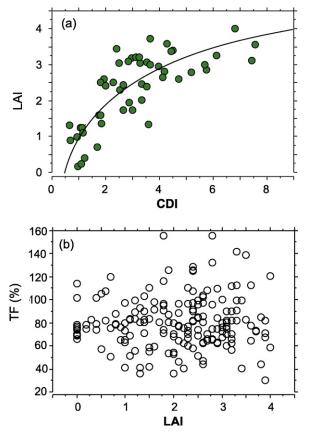


Fig. 3 – (a) Relationship between canopy density index (CDI) and leaf area index (LAI). The logarithmic regression is significant ($r_{\rm adj}^2=0.65$; p-value < 0.0001) for the 51 data pairs taken at the locations identified in Fig. 1e. (b) The plot of point throughfall (from method 3) versus LAI, both determined at the 180 measurement locations shown in Fig. 1f, reveals no relationship between the two variables. Units for LAI and CDI are $m^2 m^{-2}$.

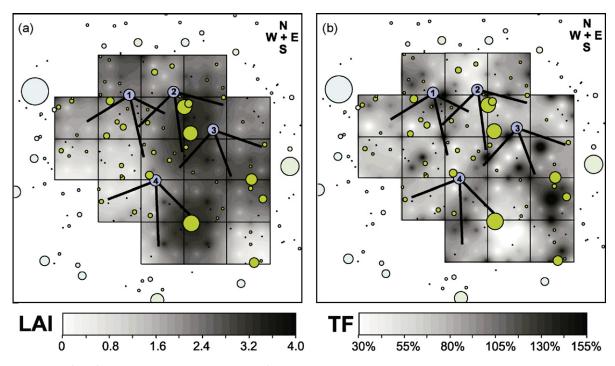


Fig. 4 – (a) LAI ($m^2 m^{-2}$) distribution within the 500 m^2 forest plot. (b) The distribution of the throughfall (TF) percentage of rainfall (determined via method 3). LAI and throughfall were interpolated as moving averages using an inverse distance weighting function (weight = 1; arbitrary limiting distance = 5 m).

occasions, rainfall depth exceeded 139 mm. Of the periods with recorded rainfall, approximately 17% had 1-min intensities >35 mm h $^{-1}$, which together accounted for 36% of the total rainfall depth. The maximum 1-min intensity was 133 mm h $^{-1}$ (9 tips min $^{-1}$).

4.3. Throughfall

Mean throughfall was 82% for both the stationary and rover methods (Table 3). Standard errors for both methods were equally low (2%); and the coefficients of variation were \leq 10%

Variable ^a	Units	Method 1	Method	Method 3-point measurements			
		Stationary gauges	2 Rovers	Entire plot	Inside stationary subplot ^a	Outside stationary subplot ^a	
Mean	%	82	82	81	82	80	
Std. dev.	%	4	8	22	25	20	
CV	%	5	10	28	31	25	
Std. error	%	2	2	2	3	2	
Min.	%	77	66	30	37	30	
Max.	%	86	95	157	157	157	
Skewness (g)	-	na ^b	-0.25	0.63	0.74	0.46	
Median	%	82	83	79	78	79	
MAD ^c	%	3	3	13	15	12	
N^{d}	_	4	20	180	73	106	
Area _{guage} e	m^2	3.08	9.66	9.66	3.94	5.72	
Area	m ²	200	500	500	200	300	

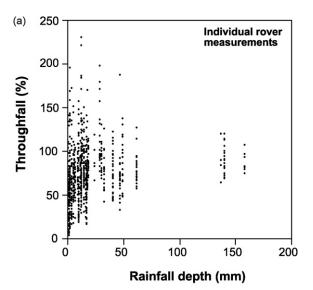
Rainfall for methods 1-2 was 1134 mm; rainfall varied by sampling occasion for the point measurements (48-177 mm).

- ^a Subplot areas shown in Fig. 1d.
- $^{\mathrm{b}}$ Not applicable because measurement number was 4.
- $^{\mathrm{c}}$ Median absolute deviation of the median.
- $^{\rm d}\,$ Number of gauges or locations for which the throughfall statistics are calculated.
- ^e Total collection area of N gauges.
- ^f Corresponding forest area sampled by the method.

(Table 3). No significant difference existed between the two methods (tied p-value = 0.47, n = 49 sampling occasions).

Despite the similarity in throughfall, the range of values was quite different: throughfall for the 20 rovers ranged from 66% to 95%; the four stationary gauges, 77–86% (Table 3). Minimum and maximum rover values during any of the 49 sampling occasions were 3% and 230%, compared with 23% and 168% for the stationary gauges (not shown). The range difference results because stationary collectors integrate measurements over an area roughly equivalent to 14 rovers: 0.77 m² versus 0.054 m² (cf. Cuartas et al., 2007).

Approximately 18% of the 835 individual rover throughfall measurements were \geq 100% (Fig. 5a). In comparison, only 7% of the 196 values recorded by the stationary gauges were \geq 100%. Stationary gauges 1–4 recorded throughfall >100% during 1, 2, 2,



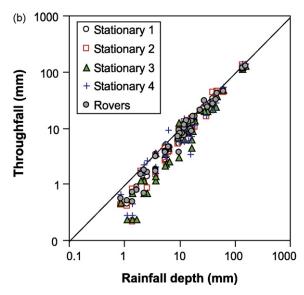


Fig. 5 – (a) Throughfall values, determined for 835 individual rover samples, demonstrate a general decrease in throughfall variability with increasing rainfall. (b) Rover- and stationary-gauge-estimated throughfall for each of the 49 sampling occasions, plotted versus associated rainfall depth.

and 8 sampling occasions, respectively. Throughfall percentage was highly variable for rainfall depths less than about 30 mm, but both rover- and stationary-based estimates increased asymptotically towards 100% for sampling occasions with large rainfall depths (Fig. 5b). Throughfall for individual stationary gauges was 81%, 83%, 77%, and 86%. No statistical difference was found between estimates made by the rovers and any group of three stationary gauges; this was not true for groups of two stationary gauges. The sampling area of three stationary gauges $(2.3~{\rm m}^2)$ was <25% the total area sampled by the rovers.

4.4. Point measurements of throughfall

Mean throughfall of the 180-point measurements was 81% (Table 3). Individual values determined from 2 to 6 sampling occasions ranged from 30% to 157% of rainfall (Table 3). Compared with the rover data of method 2, these point-based (method 3) data were significantly skewed (g=0.63 versus g-critical = ± 0.357 at $\alpha=0.05$) and the coefficient of variation was higher (28% versus 10%). No statistical difference was found between point throughfall measured inside versus outside the area sampled by the four stationary gauges (unpaired t-test, tied p-value = 0.80; area shown in Fig. 1d). Mean throughfall values for these 200 and 300 m² subplots were 82 and 80%, respectively (Table 3). Neither estimate was different from the value determined for the entire 500 m² plot (ANCOVA, $\alpha=0.05$).

About 15% of the 180-point throughfall values were \geq 100% (Fig. 4b). These canopy drip points were observed somewhere in the plot during 44 of 49 sampling occasions; however, no significant correlation existed with depth of rainfall. Furthermore, no predictive relationship existed between LAI and throughfall (Fig. 3b). Thus, the spatial structure of throughfall in the forest stand was not simply a function of the immediate overhead canopy cover. Furthermore, a sill was not reached during variogram analysis, indicating that all throughfall measurement locations in the 500 m² plot were correlated to some degree.

5. Discussion

5.1. Comparison with other SE Asian studies

The throughfall estimate obtained in this study is within the range of those reported for several forest types of various disturbance conditions in the SE Asia region (Table 4). Importantly, it is lower than the 86-91% range reported by Tanaka et al. (2001, 2005) for a montane evergreen forest at the Kog Ma research station located <50 km from PKEW. That study measured throughfall with 30 randomly located, nonroving collectors for 4 years. Vegetation characteristics were somewhat different at the two sites: e.g., Lithocaropus spp. and Quercus spp. were among the most abundant trees at Kog Ma (cf. Table 2); and most canopy trees at Kog Ma were 25-40 m in height. The main difference between the two forests was that fog precipitation totalling as much as 5-10% of rainfall augmented throughfall at Kog Ma (Tanaka, pers. commun.). Fog was not a significant precipitation component at PKEW. If fog is taken into consideration, the PKEW estimated falls within the range of values reported at Kog Ma.

Table 4 – Throughfall o	estimates for	various forests in SE
Location	Throughfall (%)	Forest type
Brunei ^a	81	Mature mixed dipterocarp rainforest
Kalimantan, Indonesia ^b	87	Lowland unlogged tropical rainforest
Peninsular Malaysia ^c	70	Lowland tropical rainforest
Peninsular Malaysia ^d	81	Primary lowland mixed dipterocarp
Peninsular Malaysia ^e	84	Tropical evergreen (kelat kendondong)
Sabah, Malaysia ^{f,g}	81–83	Lowland dipterocarp rainforest
Sabah, Malaysia ^{h,i}	91	Lowland dipterocarp
Sarawak, Malaysia ^j	82–88	Lowland evergreen tropical
Sulawasi, Indonesia ^k	70	Lower montane rainforest
Taiwan ^l	92	Moist subtropical mixed evergreen
Thailand (Kog Ma) ^m	86-91	Montane evergreen
Thailand (PKEW) ⁿ	82	Montane evergreen-dominated
^a Dykes (1997). ^b Asdak et al. (1998).		
^c Saberi and Rosnani (2004	1)	
d Konishi et al. (2006).	t).	
e Zulkifli et al. (2001).		
f Sinun et al. (1992).		
g Burghouts et al. (1998).		
h Chappell et al. (2001).		
i Bidin and Chappell (2004)	1	
^j Manfroi et al. (2006).	,.	
k Dietz et al. (2006).		
l Lin et al. (2000).		
^m Tanaka et al. (2001, 2005	5).	
1 allana Ct al. (2001, 2003	7.	

5.2. Sufficient experimentation

ⁿ This study.

The study was conducted for just 4 months, during which about 70% of a yearly total of 1600 mm of rainfall fell. By the 36th

sampling occasion, when approximately 730 of the total 1134 mm rain had fallen, throughfall estimated by the stationary gauges and rovers was indistinguishable (Fig. 6a and b). At this time, standard errors for all three-estimate methods were ≤2%—and no substantial change in mean throughfall took place afterwards (Fig. 6). Prior works have indicated that reliable throughfall estimates should have standard errors \leq 5–10% (e.g., Kimmins, 1973; Rodrigo and Avila, 2001, Holwerda et al., 2006), indicating a sufficient number of sampling occasions had been included to obtain reliable estimates in this study. One drawback to method 3 was that throughfall at each of the 180 points was based on rainfall depths (48-177 mm, Table 1) that were much smaller than those after which the throughfall variation determined by the 20 rovers of method 1 stabilized (600-700 mm; Fig. 6a and b). A better approach may have been to sample until the variation at each point had stabilized.

Fixed-gauge methods typically produce higher CVs than roving methods; and therefore, more collectors are required to estimate throughfall at the same level of confidence and error (Holwerda et al., 2006). This was observed in our study, as the CV of the point method was 28% versus 10% for rover method 2 (Fig. 6c; Table 3). Based on these values, only about 31-point measurements versus four rovers would be needed to estimate throughfall at the 95% confidence interval for a 10% error (Eq. (1)). Both of these minimum sample sizes are smaller than the total measurement points or roving gauges employed in this study.

Of interest, however, was that the rover CVs were higher than those of the stationary gauges (10% versus 5%, Table 3). Again, this results because each stationary gauge integrates throughfall over a larger area. Importantly, reliable throughfall estimates with these gauges were facilitated by dynamic calibration of the tipping-bucket mechanism. Had we assumed a constant depth-per-tip relationship, individual stationary gauge throughfall estimates would have been underestimated by 20%; total rainfall depth would have been underestimated by about 15%.

5.3. Spatial structure

Some drip points occurred near crown perimeters and the trunks of large canopy trees (Fig. 7). Throughfall in the vicinity

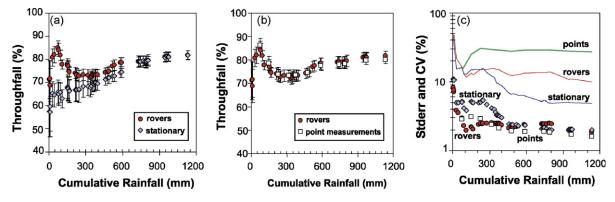


Fig. 6 – (a) Stabilization of the throughfall estimate, determined by the stationary (method 1) and rover (method 2) gauges. (b) Stabilization of the throughfall estimates, determined by rover methods 2 and 3 (point measurements). The values in panels (a) and (b) are mean \pm standard errors. (c) For all three methods, the change in standard errors (symbols) and coefficients of variation (curves) are plotted for accumulating rainfall.

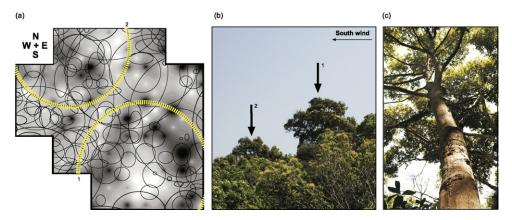


Fig. 7 – (a) Idealized representation of the crown perimeters for all trees affecting throughfall in the study plot. Interpolated throughfall is shown in the background (cf. Fig. 4b). The crown perimeters of the two trees shown in panel (b) are hashed. (b) Number 1 indicates the location of the 35-m tall Eugenia albiflora tree on the SE corner of the plot that potentially prevents rainfall from entering the plot during periods of strong south winds that produce non-vertical rain-driven wind. Number 2 marks the location of an upper-story Castanopsis tribuloides tree that creates a canopy drip point (panel c). (c) Canopy and trunk of a C. tribuloides (#2, panel b) for which stemflow water traveling down the trunk frequently exists as throughfall at many of the knots shown.

of a handful of upper-story Castanopsis tribuloides trees was often substantial because stemflow originating high in the sparse canopy flowed in a concentrated stream to common exit points at knots on the tree trunk or large branches (Fig. 7c). Our sampling strategy, however, did not reveal an obvious spatial structure in throughfall, as has been found by others (e.g., Whelan and Anderson, 1996; Llorens and Gallart, 2000). Similarly, Loescher et al. (2002) reported only a weak relationship between canopy cover and throughfall volume (cf., Bellot and Escarre, 1991); and Carlyle-Moses et al. (2004) determined that throughfall was only related to various upper- and understory canopy variables during relatively small rainfall events. The inability to ensure that estimated LAI values represented the actual canopy area affecting throughfall at any given point may have hampered our ability to identify such a relationship (Levia and Frost, 2006).

Konishi et al. (2006) found cyclic patterns in the spatial distribution of throughfall for distances greater than about 11 m in their study plot in Peninsular Malaysia. Others have noted that throughfall gauges should be separated by anywhere from half a crown diameter to >40 m to be spatially independent (Brouwer, 1996; Loescher et al., 2002; Keim et al., 2005). Given the lack of independence among throughfall measurement locations, it is important to be cautious when using these small-plot results to 'scale up' to the watershed level (cf. Scatena, 1990; Chappell et al., 2001; Manfroi et al., 2006).

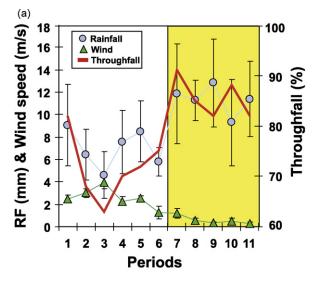
Spatial structure in throughfall in PKEW might have been affected by the interaction of topography, variable tree height, and wind-driven rain (Herwitz and Slye, 1992, 1995). For example, low throughfall observed during the first six periods of the study coincided with comparatively strong southerly winds (Table 1, Fig. 8). Strong winds can produce substantially non-vertical rainfall in PKEW. In the case of southerly winds, the canopy of a 35-m tall Eugenia albiflora tree located on the SE corner of the plot would likely block non-vertical rainfall from

entering the forest plot (Fig. 7a and b). An alternative explanation is an increase in canopy cover over the course of the study. However, the few deciduous trees within the stand were fully leafed by the beginning of the experiment; and no noticeable change in canopy vegetation density occurred thereafter.

5.4. Practicality of stationary gauges

Reduction of the number of gauges required for estimating throughfall with high statistical confidence and low error is desirable for minimizing cost and effort (Helvey and Patric, 1965; Rodrigo and Avila, 2001; Holwerda et al., 2006). The stationary gauges, which produced a throughfall estimate indistinguishable from the rovers, are advantageous because they integrate a sampling area equivalent to several rovers, do not require daily visits because they record data automatically, and do not have to be periodically repositioned. The stationary gauges also allow detailed examination of the rainfall interception process over the course of individual storms (Link et al., 2004; Cuartas et al., 2007). Furthermore, in remote and difficult terrain, reliable automatic recording gauges may be the only practical option for obtaining realistic throughfall estimates when determining the spatial distribution is not a goal.

Drawbacks include disruption of the understory in the immediate area around the troughs and routine site visits to clear the troughs of debris and repair them if damaged by limb-fall. The stationary gauge method is also substantially more expensive that the rover setup, because of the cost of fabrication, and the necessity of using a data logger and portable computer for downloading. Furthermore, the large volumetric capacity of the tipping mechanism allows for substantial error for small throughfall depths; and dynamic calibration of the tipping-bucket mechanism can be tedious in the field.



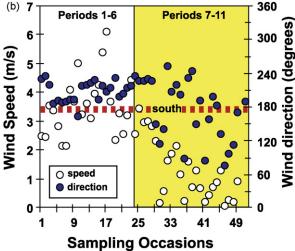


Fig. 8 – (a) Mean wind speed, rainfall, and throughfall for each of the 11 periods during the investigation. Error bars are standard errors. (b) Mean wind speed and direction during the 49 sampling occasions. Compared with the latter periods, sampling occasions during the initial periods were characterized by stronger south-dominated winds and lower throughfall.

6. Conclusions

Identical throughfall estimates (82%) in the 500-m² disturbed evergreen-dominated montane forest plot in northern Thailand were determined with mobile and stationary trough methods—but only after sampling more than about 700 mm, collected during 36 sampling occasions. The range of throughfall depths recorded during 49 sampling occasions varied between methods because the stationary gauges integrated throughfall over a much larger collection area than individual rovers. Two important keys for obtaining reasonable throughfall estimates were correcting for splash loss errors among the throughfall and rainfall collection gauges and performing dynamic calibration on all tipping-bucket mechanisms.

Standard errors (2%) and coefficients of variation (\leq 10%) for both mobile and stationary methods were low, in part, because of the absence of tree gaps. While we found no correlation between leaf area index and throughfall at corresponding understory locations, some canopy drip points (i.e., areas with throughfall greater than 100%) were located near crown perimeters of canopy trees and near the base of certain upper-story trees that generated substantial stemflow which dripped to the ground from the trunks and large branches. However, all throughfall measurement locations in the small patch were correlated to some degree, possibly because of the interaction of wind-driven rainfall and one dominant canopy tree. Although the throughfall estimates were similar to those observed at a nearby forest in northern Thailand, they might not be an accurate predictor of throughfall at the watershed scale in the study area, because of the lack of canopy gaps in the experimental plot and the variable topography in the basin.

Acknowledgments

Partial funding for this project was received from the National Science Foundation USA (Grants 9614259, 0000546). ADZ was supported by an Environmental Protection Agency Star Fellowship. The research was sanctioned by National Research Council Thailand (NRCT) and the Thai Royal Forest Department. Tree identification was performed by J.F. Maxwell (Chiang Mai). Additional measurement assistance was provided by Kalaya Kantawong.

REFERENCES

Aboal, J.R., Jiménez, M.S., Morales, D., Gil, P., 2000. Effects of thinning on throughfall in Canary Islands pine forest—the role of fog. Journal of Hydrology 238, 218–230.

Asdak, C., Jarvis, P.G., van Gardingen, P., Fraser, A., 1998. Rainfall interception loss in unlogged and logged forest areas of Central Kalimantan, Indonesia. Journal of Hydrology 206, 237–244.

Bellot, J., Escarre, A., 1991. Chemical characteristics and temporal variations of nutrients in throughfall and stemflow of three species of Mediterranean holm oak forest. Forest Ecology and Management 41, 125–135.

Bidin, K., Chappell, N.A., 2004. Sub-canopy rainfall and wetcanopy evaporation in a selectively-logged rainforest, Sabah, Malaysia. Water: forestry and land use perspectives. In: Abdul Rahim, N. (Ed.), IHP-VI, Technical Documents in Hydrology, No. 70. UNESCO, Paris, pp. 69–85.

Bouten, W., Heimovaara, T.J., Tiktak, A., 1992. Spatial patterns of throughfall and soil water dynamics in a Douglas fir stand. Water Resources Research 28, 3227–3233.

Brouwer, L.C., 1996. Nutrient cycling in pristine and logged tropical rain forest. Tropenbos-Guyana series 1. Tropenbos-Guyana Programme, Guyanna, Georgetown.

Bruijnzeel, L.A., 1990. Hydrology of moist tropical forests and effects of conversion: a state of knowledge review. Humid Tropics Programme of the International Hydrological Programme of UNESCO Paris, and Vrije Universiteit Amsterdam

Bruijnzeel, L.A., 2000. Forest hydrology. In: Evans, J.C. (Ed.), The Forests Handbook. Blackwell, Oxford, pp. 301–343, Chapter 12.

- Bruijnzeel, L.A., Wiersum, K.F., 1987. Rainfall interception by a young Acacia auriculiformis A. Cunn. Plantation forest in West Java, Indonesia: application of Gash's analytical model. Hydrological Processes 1, 309–319.
- Burghouts, T.B.A., Van Straalen, N.M., Bruijnzeel, L.A., 1998. Spatial heterogeneity of element and litter turnover in a Bornean rain forest. Journal of Tropical Ecology 114, 477–506.
- Calder, I.R., 2001. Canopy processes: implications for transpiration, interception and splash induced erosion, ultimately for forest management and water resources. Plant Ecology 153, 203–214.
- Calder, I.R., Kidd, C.H.R., 1978. A note on the dynamic calibration of tipping-bucket gauges. Journal of Hydrology 39, 383–386
- Calder, I.R., Rosier, P.T.W., 1976. The design of large plastic sheet net rainfall gauges. Journal of Hydrology 30, 403–405.
- Carlyle-Moses, D.E., Fores Laureano, J.S., Price, A., 2004.

 Throughfall and throughfall spatial variabiltiy in Madrean oak forest communities of northeastern Mexico. Journal of Hydrology 297, 124–135.
- Chappell, N.A., Bidin, K., Tych, W., 2001. Modelling rainfall and canopy controls on net precipitation beneath selectivelylogged tropical forest. Plant Ecology 153, 215–229.
- Crockford, R.H., Richardson, D.P., 2000. Partitioning of rainfall into throughfall, stemflow and interception: effect of forest type, ground cover and climate. Hydrological Processes 14, 2903–2920.
- Cuartas, L.A., Tomasella, J., Nobre, A.D., Hodnett, M.G., Waterloo, M.J., Munera, J.C., 2007. Interception waterpartitioning dynamics for a pristine rainforest in Central Amazonia: marked differences between normal and dry years. Agriculture and Forest Meteorology 145, 69–83.
- Czarnowski, M.S., Olszewski, J.L., 1970. Number and spacing of rainfall gauges in a deciduous forest stand. Oikos 21, 48–51.
- Davis, J.C., 2002. Statistics and Data Analysis in Geology, 2nd ed. John Wiley & Sons, Inc., New York.
- Decagon Services, 2004. AccuPAR PAR/LAI Ceptometer Model LP-80, Operator's Manual Version 1.2. Decagon Devices, Inc., Pullman, WA 99163. (downloaded June 2005 from www.decagon.com/lp80).
- Dietz, J., Hölscher, D., Leuschner, C., Hendrayanto, 2006. Rainfall partitioning in relation to forest in Central Sulawesi, Indonesia. Forest Ecology and Management 237, 170–178.
- Dingman, S.L., 1994. Physical Hydrology. Prentice-Hall, Inc., Upper Saddle River, NJ.
- Dunkerley, D., 2000. Measuring interception loss and canopy storage in dryland vegetation: a brief review and evaluation of available research strategies. Hydrological Processes 14, 669–678.
- Dykes, A.P., 1997. Rainfall interception from a lowland tropical rainforest in Brunei. Journal of Hydrology 200, 260–279.
- Forti, M.C., Neal, C., 1992. Spatial variability of throughfall chemistry in a tropical rainforest (Central Amazonia, Brazil). Science of the Total Environment 120, 245–259.
- Germer, S., Elsenbeer, H., Moraes, J.M., 2006. Throughfall and temporal trends of rainfall redistribution in an open tropical rainforest, south-western Amazonia (Rondonia, Brazil). Hydrology and Earth System Sciences 10, 383–393.
- Hall, R.L., 2003. Interception loss as a function of rainfall and forest types: stochastic modelling for tropical forest canopies revisited. Journal of Hydrology 280, 1–12.
- Hall, R.L., Roberts, J.M., 1990. Hydrological aspect of new broadleaf plantations. SEESOIL 6, 2–38.
- Helvey, J.D., Patric, J.H., 1965. Canopy and litter interception of rainfall by hardwoods of eastern United States. Water Resources Research 1, 193–206.
- Herwitz, S.R., Slye, R.E., 1992. Spatial variability in the interception of inclined rainfall by a tropical rainforest canopy. Selbyana 13, 62–71.

- Herwitz, S.R., Slye, R.E., 1995. Three-dimensional modeling of canopy tree interception of wind-driven rainfall. Journal of Hydrology 168, 205–226.
- Hewlett, J.D., 1982. Principles of Forest Hydrology. University of Georgia Press, Athens, GA.
- Hölscher, D., Köhler, L., Leuschner, C., Kappelle, M., 2003. Nutrient fluxes in stemflow and throughfall in three successional stages of an upper montane rain forest in Costa Rica. Journal of Tropical Ecology 19, 557–565.
- Holwerda, F., Scatena, F., Bruijnzeel, L.A., 2006. Throughfall in a Puerto Rican lower Montane rain forest: a comparison of sampling strategies. Journal of Hydrology 327, 592–602.
- Huber, A., Iroumé, A., 2001. Variability of annual rainfall partitioning for different sites and forest covers in Chile. Journal of Hydrology 248, 78–92.
- Humphrey, M.D., Istok, J.D., Lee, J.Y., Hevesi, J.A., Flint, A.L., 1997. A new method for automated dynamic calibration of tipping-bucket rain gauge. Journal of Atmospheric and Oceanic Technology 14, 1513–1519.
- Keim, R.F., Skaugset, A.E., Weiler, M., 2005. Temporal persistence of spatial patterns in throughfall. Journal of Hydrology 314, 263–274.
- Kimmins, J.P., 1973. Some statistical aspects of sampling throughfall precipitation in nutrient cycling studies in British Columbian coastal forests. Ecology 54, 1008–1019.
- Konishi, S., Tani, M., Kosugi, Y., Tankanashi, S., Sahat, M.M., Abdul Rahim, N., Niiyama, K., Okuda, T., 2006. Characteristics of spatial distribution of throughfall in a lowland tropical rainforest, Peninsular Malaysia. Forest Ecology and Management 224, 19–25.
- Kostelnik, K.M., Lynch, J.A., Grimm, J.W., Corbett, E.S., 1989. Sample size requirements for estimation of throughfall chemistry beneath a mixed hardwood forest. Journal of Environmental Quality 18, 274–280.
- Levia, D.F., Frost, E.E., 2003. A review and evaluation of stemflow literature in the hydrologic and biogeochemical cycles of forested and agricultural ecosystems. Journal of Hydrology 274. 1–29.
- Levia, D.F., Frost, E.E., 2006. Variability of throughfall volume and solute inputs in wooded ecosytems. Progress in Physical Geography 30, 605–632.
- Lin, T.C., Hamburg, S.P., King, H.B., Hsia, Y.J., 2000. Throughfall patterns in a subtropical rain forest of northeastern Taiwan. Journal of Environmental Quality 29, 1186–1193.
- Link, T.E., Unsworth, M., Marks, D., 2004. The dynamics of rainfall interception by a seasonal temperate rainforest. Agricultural and Forest Meteorology 124, 171–191.
- Llorens, P., Gallart, F., 2000. A simplified method for forest water storage capacity measurement. Journal of Hydrology 240, 131–144
- Lloyd, C.R., de O. Marques, F.A., 1988. Spatial variability of throughfall and stemflow measurements in Amazonian rainforest. Agriculture and Forest Meteorology 42, 63–73.
- Loescher, H.W., Powers, J.S., Oberbauer, S.F., 2002. Spatial variation of throughfall volume in an old-growth tropical wet forest, Costa Rica. Journal of Tropical Ecology 18, 397–407.
- Manfroi, O.J., Kuraji, K., Suzuki, M., Tanaka, N., Kume, T., Nakagawa, M., Kumagai, T., Nakashizuka, T., 2006.

 Comparison of conventionally observed interception evaporation in a 100 m² subplot with that estimated in a 4-ha area of the same Bornean lowland tropical forest. Journal of Hydrology 239, 329–349.
- Marsalek, J., 1981. Calibration of the tipping-bucket raingage. Journal of Hydrology 53, 343–354.
- Peterson, D.L., Rolfe, G.L., 1979. Determining sample size in throughfall studies. Forest Science 25, 582–584.
- Raat, K.J., Draaijers, G.P.J., Schaap, M.G., Tietema, A., Verstraten, J.M., 2002. Spatial variability of throughfall water and

- chemistry and forest floor water content in a Douglas fir forest stand. Hydrology and Earth System Sciences 6, 363–374
- Robson, A.J., Neal, C., Ryland, G.P., Harrow, M., 1994. Spatial variations in throughfall chemistry at the small plot scale. Journal of Hydrology 158, 107–122.
- Rodrigo, A., Avila, A., 2001. Influence of sampling size in the estimation of mean throughfall in two Mediterranean holm oak. Journal of Hydrology 243, 216–227.
- Saberi, O., Rosnani, H.M., 2004. Rainfall interception by lowland tropical rainforest in Air Hitam, Selangor, Malaysia. Water: forestry and land use perspectives. In: Abdul Rahim, N. (Ed.), IHP-VI, Technical Documents in Hydrology, No. 70. UNESCO, Paris, pp. 57–68.
- Scatena, F.N., 1990. Watershed scale rainfall interception on two forested watersheds in the Luquillo Mountains of Puerto Rico. Journal of Hydrology 113, 89–102.
- Schroth, G., Elias, M.E.A., Seixas, R., Zech, W., 2001. Nutrient fluxes in rainfall, throughfall, and stemflow in tree-based land use systems and spontaneous tree vegetation of central Amazonia. Agriculture, Ecosystems and Environment 87, 37–49.
- Schume, H., Jost, G., Katzensteiner, K., 2003. Spatio-temporal analysis of the soil water content in a mixed Norway spruce (Picea abies (L.) Karst.)—European beech (Fagus sylvatica L.) stand. Geoderma 112, 273–287.
- Sinun, W., Meng, W.W., Douglas, I., Spencer, T., 1992. Throughfall, stemflow, overland flow and throughflow in the Ulu Segama rain forest, Sabah, Malaysia. Philosophical Transactions of the Royal Society of London B 233, 321–346.
- Staelens, J., De Schrijver, A.D., Verheyen, K., Verhoest, N.E.C., 2006. Spatial variability and temporal stability of throughfall water under a dominant beech (Fagus sylvatica L.) tree in relationship to canopy cover. Journal of Hydrology 330, 651–662.

- Stogsdill, W.R., Wittwer, R.F., Hennessey, T.C., Dougherty, P.M., 1989. Relationship between throughfall and stand density in a Pinus taeda plantation. Forest Ecology and Management 29, 105–113.
- Stout, B.B., McMahon, R.J., 1961. Throughfall variation under tree crowns. Journal of Geophysical Research 66, 1839–1843.
- Tanaka, N., Tantasirin, C., Suzuki, M., Tangtham, N., 2001.
 Rainfall interception by a hill evergreen forest in northern
 Thailand. In: The Fifth International Study Conference on
 GEWEX in Asia and GAME. Proceedings of the GAME
 International Science Panel, vol. 2. Hydrospheric
 Atmospheric Research Center, pp. 410–414.
- Tanaka, N., Tantasirin, C., Kuraji, K., Suzuki, M., Tangtham, N., 2005. Inter-annual variation in rainfall interception at a hill evergreen forest in northern Thailand. Bulletin of Tokyo University Forestry 113, 11–44.
- Whelan, M.J., Anderson, J.M., 1996. Modelling spatial patterns of throughfall and interception loss in a Norway spruce (Picea abies) plantation at the plot scale. Journal of Hydrology 186, 335–354.
- Zar, J.H., 1999. Biostatistical Analysis, 4th ed. Prentice Hall, Upper Saddle River, NJ.
- Ziegler, A.D., Giambelluca, T.W., Sutherland, R.A., Nullet, M.A., Yarnasarn, S., Pinthong, J., Preechapanya, P., Jaiarree, S., 2004. Toward understanding the cumulative impacts of roads in agricultural satersheds of montane mainland southeast Asia. Agriculture, Ecosystems and Environment 104, 145–158.
- Zulkifli, Y., Abdul Rahim, N., Baharuddin, K., 2001. Rainfall partitioning in an old growth tropical rain forest. In: Mohamed, A.H., et al. (Eds.), Tropical Forestry Research in the New Millennium: Meeting Demands and Challenges. Proceedings of the International Conference on Forestry and Forest Products Research (CFFPR 2001), Kuala Lumpar, Malaysia, October 1–3, 2001, pp. 1–9.